Photodegradation of methyl red by advanced and homogeneous photo-Fenton's processes: A comparative study and kinetic approach

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The degradation of methyl red (MR), an azo dye, was carried out by the homogeneous photo-Fenton's process (HPFP) and the advanced photo-Fenton's process (APFP) using symmetrical peroxides such as hydrogen peroxide and ammonium persulfate (APS) as oxidants. The APFP showed higher efficiency than their homogeneous counterparts even at high dye concentrations due to the faster reduction of Fe^{3+} to Fe^{2+} ions on the iron surface. H_2O_2 proved to be a better oxidant for both the processes. However, APS efficiently inhibited the precipitation of iron oxy hydroxides at higher dosage of iron powder compared to H_2O_2 by providing excess acidity to the reaction medium. The rate constant for the kinetics of decolorisation by various oxidation processes is of the order: $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV} > \text{Fe}^0/\text{H}_2\text{O}_2/\text{dark} > \text{Fe}^0/\text{APS}/\text{UV} > \text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV} > \text{Fe}^0/\text{UV} > \text{Fe}^0/\text{APS}/\text{dark} > \text{Fe}^0/\text{dark} \approx \text{H}_2\text{O}_2/\text{UV} > \text{Fe}^{2+}/\text{APS}/\text{UV} > \text{APS}/\text{UV} > \text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{dark} > \text{Fe}^{2+}/\text{APS}/\text{dark} \approx \text{Fe}^{2+}/\text{UV}$. The degradation reaction was followed by UV-visible and GC-MS spectroscopic techniques. Based on the intermediates obtained, probable degradation mechanisms have been proposed. It was found that the initial mechanism in the APFP involves the reduction of azo groups to amines while in the case of HPFP it leads to the formation of hydroxylated products due to the oxidation of azo groups.

1. Introduction

Azo dyes contain one or more nitrogen to nitrogen double bond ($-\text{N}=\text{N}-$) and they constitute almost 80% of the reactive dyes. Most of the azo dyes are not biodegradable by aerobic treatment processes.¹ However, azo dyes can be decolorised by anaerobic treatment,^{2,3} but this leads to the production of toxic by-products such as aromatic amines.⁴ Advanced oxidation processes involving Fenton's and photo-Fenton's process for the degradation of organic dyes are extensively reported in the literature.^{5–10} The main advantage of photo-Fenton's process are; reagents are easily available, cheap and cost effective. Further no scattering of light takes place, contact between the pollutants and the catalyst is high since the reaction is homogeneous. The disadvantage of a homogeneous process is the non-reusability of the catalyst which hinders its commercial application. Therefore, many researchers have investigated the applications of iron incorporated zeolites¹¹ and iron minerals¹² as catalytic material for the oxidation of organic pollutants. Lucking *et al.*¹³ tested the catalytic properties of iron powder, graphite and activated carbon for the oxidation of 4-chlorophenol in aqueous solution with hydrogen peroxide. They concluded that iron powder could be used to replace iron salts as a catalyst to the Fenton's process. Zero Valent Metallic Iron (ZVMI) powder is an environmental friendly reducing agent which has been widely used for environmental remediation in recent years. In the present study, we compare the efficiency of the homogeneous photo-Fenton process (HPFP) with the Advanced photo-Fenton process

(APFP) for the degradation of the azo dye—methyl red (MR). Though H_2O_2 is extensively used in Fenton's process, few attempts have been made towards the use of peroxy disulfate as an oxidant. In view of this, the main focus is on the utilization of symmetrical peroxides like hydrogen peroxide and ammonium persulfate (APS) as the oxidants in the Fenton's and photo-Fenton's process.

2. Experimental section

2.1 Materials and methods

MR, methyl alcohol, APS, ferrous oxalate (FeC_2O_4), iron powder (electrolytic, 300 mesh size, 95% purity) and hydrogen peroxide (50%) were obtained from S. D. Fine Chemicals, Bombay, India and were used as received. The structure of the dye is shown in Scheme 1. The molecular formula of MR is $\text{C}_{15}\text{H}_{15}\text{N}_3\text{O}_2$ and the molecular weight is 269.

2.2 Experimental procedure

All the experiments were performed in a glass reactor of one litre capacity whose surface area is 176 cm^2 . The light source used was a medium pressure mercury vapor lamp of 125 W. The photon flux of the light source was found to be 7.75 mW cm^{-2} as determined by ferrioxalate actinometry and the wave length emission is in the range of 350–400 nm. The experiments were performed in the presence of atmospheric oxygen using distilled water. In a typical experiment of HPFP, 200 ml of 10 ppm dye solution was taken along with iron ion solution and was exposed directly to the light. The pH of the solution reduces from 6.6 to 2.3 and it was adjusted to 3 using dilute NaOH. In APFP, pH of the dye solution after the addition of iron powder was found to

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Scheme 1 Probable degradation pathway for MR by APFP.

be 6.6 and it was adjusted to 3 by adding dilute H₂SO₄. The oxidant of desired concentration was added in both the process at the beginning of the experiment.

2.3 Analytical methods

The solution was taken out from the reactor at definite time intervals and was centrifuged. The centrifugates were analyzed by UV-visible spectroscopy using Shimadzu UV-1700 Pharmaspec UV-visible spectrophotometer. The concentration of the dye was calculated from the standard calibration curve at its characteristic λ_{max} of MR dye. The centrifugates were extracted into non-aqueous medium and were subjected to GC-MS analysis using GC-MS-QP 5000 Shimadzu mass spectrometer to identify the photo products during the course of the reaction.

3. Results and discussion

3.1 The various possible reactions taking place at the iron surface/iron ions with the oxidants are illustrated as follows

Iron surface gets oxidized in acidic medium generating ferrous ions (Fe²⁺).

Alternatively these ions can also be generated by the reaction of iron surface with H₂O₂. Fe²⁺ ions so formed are partially adsorbed on the iron surface. These adsorbed Fe²⁺ ions can react with H₂O₂ leading to the generation of hydroxyl radicals.

If these Fe²⁺ ions are not partially adsorbed, they can diffuse into the bulk of the solution generating hydroxyl anions. Fe⁰ initially oxidizes to Fe²⁺ by losing two electrons as shown in eqn (4). Two electrons are transferred to electron acceptors like H₂O₂ which can form two hydroxide ions as shown in eqn (5). On adding eqn (4) and (5) we get eqn (6).

The oxidant APS can react with iron surface forming sulfate and ferrous ions.

The Fe^{2+} ions generated in the above reactions can further react with the respective oxidant molecules generating free radicals and simultaneously getting oxidized to form Fe^{3+} ions.

In eqn (3), Fe^{2+} ions formed on the solid iron surface can react with the oxidizing agent whose rate constant is $6 \times 10^{-2} \text{ M}^{-1} \text{ s}^{-1}$, while in eqn (8) it is the free Fe^{2+} ions in the solution that react with the oxidants which shows a rate constant of $76 \text{ M}^{-1} \text{ s}^{-1}$. Though the nature of the products remains same in both the cases, the rate of formation of Fe^{3+} ions is faster in the homogeneous condition [eqn (9)].¹⁴

Ferric ions (Fe^{3+}) so formed can either react with molecules of water or H_2O_2 as represented in the following way:

(i) Ferric ions (Fe^{3+}) on reaction with water form an hydroxo complex. This complex under UV-irradiation generates Fe^{2+} ions and hydroxyl radicals.

(ii) Fe^{3+} ions can also react with H_2O_2 leading to the formation of Fe^{2+} ions along with hydroperoxyl radicals. Further, so formed hydroperoxyl radicals can reduce Fe^{3+} ions, along with the generation of hydroxyl radicals. The Fe^{2+} ions are thus formed in two stages which can actively participate in the cyclic Fenton's process.

Two molecules of sulfate radicals are produced when Fe^{3+} reacts with persulfate anions, which can further react with water molecule generating hydroxyl radicals.

The efficiency of the photo-Fenton process was compared with Fenton's process (dark). Only 12 and 32% of the dye was decolorised for the process $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{dark}$ and $\text{Fe}^0/\text{H}_2\text{O}_2/\text{dark}$, respectively, whereas complete decolorisation of the dye takes around 90 and 15 min for the above mentioned process, respectively, under UV illumination. This is due to the fact that back reduction process of Fe^{3+} to Fe^{2+} ions, is a rate determining step in dark which proceeds at a faster rate in the presence of UV light. The faster decolorisation in the APFP is attributed to the excess generation of Fe^{2+} ions by the iron surface which reacts with H_2O_2 generating higher numbers of hydroxyl radicals enhancing the decolorisation rate.

3.2 Effect of pH

To find the optimum pH, experiments of $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$ and $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$ were carried out at different pH values in the range of 1.02 to 9.5 by maintaining the other reaction parameters constant. In the pH range 2.5 to 3.0, complete decolorisation takes place for both the processes $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$ and $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$ at 15 and 90 min respectively. This is because at pH 3, it is proposed that half of the iron species exist as Fe^{3+} ions and half as a complex of $\text{Fe}[\text{OH}]^{2+}(\text{H}_2\text{O})_5$ which are the dominant photo active species.^{8,15-17} They have the largest light absorption coefficient and show high quantum yield for hydroxyl radical generation, along with regeneration of Fe^{2+} ions in the wavelength range of 280–370 nm.¹⁸ However, the exact ratio of these photoactive species is uncertain according to Baes and Mesmer¹⁹ and also by Barnum²⁰ because the measurements generally suffer from uncertainty.

Beyond this optimum pH, the degradation rate decreases slowly and in alkaline pH (>7) the dye resists degradation. The decrease in the decolorisation rate at lower pH (<2) values can be attributed to the presence of excess protons acting as hydroxyl radical scavengers according to eqn (16).

The lower efficiency at higher pH values (>7) is due to the precipitation of iron as $\text{Fe}(\text{OH})_3$ which gets deposited on the iron surface in APFP and remains in the solution in HPFP thus blocking the transparency of the solution. This prevents the electron transfer process between Fe^0 and oxidants in the APFP process thereby reducing the concentration of hydroxyl radicals in the solution. Therefore both HPFP and APFP were optimized at pH 3.

3.3 Effect of oxidants

Voluminous literature can be found for the use of H_2O_2 as the oxidant in the Fenton's/photo-Fenton's process.⁵⁻¹⁰ The application of peroxy disulfate as an oxidant in the Fenton's process is least explored. Persulfate anion is the most powerful oxidant of the peroxygen family of compounds and one of the strongest oxidants used in remediation.

The standard oxidation–reduction potential for the above reaction is 2.1 V compared to 1.8 V for H_2O_2 and 1.4 V for peroxy monosulfate anion (HSO_5^-). This potential is higher than the redox potential of the permanganate anion (1.7 V), but slightly lower than that of ozone (2.2 V). In addition to its direct oxidation capacity, persulfates can be induced to form sulfate radicals, thereby providing free radical reaction mechanisms similar to the hydroxyl radical pathways generated by Fenton's chemistry.

The sulfate radical is one of the strong aqueous oxidizing species with a redox potential of 2.6 V similar to that of the

hydroxyl radical 2.8 V. The sulfate radicals are then involved in a series of reactions leading to the degradation of organic compounds. The sulfate radicals produced engages in at least three reaction modes with organic compounds: (i) by abstracting a hydrogen atom from saturated carbon, (ii) by adding to unsaturated or aromatic carbon, (iii) by removing one electron from carboxylate anions and from certain neutral molecules.²¹ These attributes make persulfate a viable option for the treatment of a broad range of organic contaminants in the aqueous medium.

When the concentration of H_2O_2 was increased from 2 to 10 ppm, the rate constant increased for both APFP and HPFP. The observed increase in the rate constant was greater for APFP compared to HPFP (Fig. 1). Beyond this concentration, the rate constant decreased for both the process and the role of oxidants ($\text{H}_2\text{O}_2/\text{APS}$) in the rate-controlling steps of the process is negligible especially at higher concentrations. At higher H_2O_2 concentrations, the excess hydroxyl radicals generated may undergo recombination reactions or it may take part in unwanted reaction pathways. Similar variations were observed for APS as an oxidant and APS was optimized at 20 ppm for APFP and 30 ppm for HPFP (Fig. 2). The reaction proceeds at a faster rate for H_2O_2 as an oxidant compared to APS for both HPFP and APFP. But in the case of APFP the extent of increase in the rate constant is almost 2.5 times higher for H_2O_2 compared to APS. The reactivity of APS on the surface of iron powder is slow compared to H_2O_2 which accounts for the observed low rate constant for APS. Hydroxyl radicals generated from H_2O_2 are either due to direct photolysis or from the reaction of H_2O_2 on the iron surface which is a single-step process. But in the case of APS, persulfate radicals on reaction with iron surface initially leads to the formation of sulfate radicals. These sulfate radicals on subsequent reaction with water molecules generate hydroxyl radicals which is two step processes. The kinetics of the later reaction is slow and due to sulfate radicals being more acidic in nature lowers the pH of the reaction medium. The final pH of the reaction medium for H_2O_2 as an oxidant was found to be 3.2–3.3 for both HPFP and APFP. The final pH of the reaction medium for HPFP using APS varied from 2.7–2.9 for all the concentrations of APS (10 to 50 ppm) while for APFP the pH of the reaction medium drastically reduced from 2.2–2.3 for higher

Fig. 1 Effect of H_2O_2 on APFP and HPFP.

Fig. 2 Effect of APS on APFP and HPFP.

concentration of APS (50 ppm). This decrease in the pH increases the concentration of protons in the solution acting as hydroxyl radical scavengers. Hence the rate constant drastically decreased to lower values for APFP compared to HPFP. This clearly suggests that the excess protons generated by APS exert negative effects on the degradation rate. Hence it can be concluded that APS effectively lowers the pH of the reaction medium for APFP compared to HPFP.

3.4 Effect of iron concentration

Experiments were performed to study the variation in the degradation efficiency at different catalyst concentrations for both HPFP and APFP. Iron ion concentration is a very important parameter in the Fenton's process as it influences the final pH of the solution and it also influences the extent of generation of hydroxyl radicals. In the case of APFP, when the concentration of iron powder was increased from 10 to 30 mg, the dye was completely decolorised within 15 and 30 min of UV illumination for H_2O_2 and APS as oxidants respectively. Beyond this iron concentration, the rate of degradation decreased due to the non-availability of a sufficient number of dye molecules and the role of oxidants ($\text{H}_2\text{O}_2/\text{APS}$) in the rate-controlling steps of the process is negligible. At high doses of iron, the excess iron might precipitate as iron oxy hydroxide or it can compete for hydroxyl radicals along with the dye molecules and act as an hydroxyl radical scavenger.²²

Hence the catalyst concentration was optimized at 10 mg/200 ml (50 mg L^{-1}) in APFP. Similar experiments were performed for HPFP and the catalyst concentration was optimized at 15 mg/200 ml (75 mg L^{-1}) for both the oxidants. The effect of iron concentration on the degradation rate was significantly influenced by the nature of the oxidants used which can be explained as follows: when iron dosage was increased from 10 to 30 mg using H_2O_2 , the final pH of the solution was 3.92 and the solution was turbid for APFP. With an increase in the iron dosage from 50 to 150 mg, the final pH of the solution varied from 5.34–5.42 and the precipitate of iron oxy hydroxides coagulates. Further, the precipitates deposit on the iron surface thus blocking the

process of electron transfer from catalyst surface to the dye molecules and also to the oxidizing agent. This reduces the concentration of hydroxyl radicals in the solution thereby decreasing the degradation efficiency. In the case of APS, the final pH of the solution was found to be 4.42–4.48 for 150 mg of iron powder. At these conditions precipitation of iron was not observed. Hence it can be concluded that APS effectively inhibits the precipitation of excess iron by providing excess acidity. However in the case of HPFP, iron precipitation was not observed at higher concentrations probably due to the homogeneous condition of the reaction medium.

3.5 Effect of substrate concentration

The degradation efficiency decreases when the concentration of dye is increased. This is due to the insufficient availability of hydroxyl radicals as the other reaction parameters like catalyst/oxidant concentration, intensity of the light and irradiation time are held constant. Further, with increase in dye concentration, most of the UV photons are absorbed by the dye molecules rather than the catalyst and oxidizing agent which further reduces the generation of hydroxyl radicals in the solution^{23,24} and leads to a photo sensitization process. APFP retains its efficiency to degrade MR even at higher concentrations (20 ppm), while HPFP shows its efficiency only at lower concentrations (10 ppm) beyond which its efficiency decreases drastically. This is attributed to the faster regeneration of Fe^{2+} ions on the iron surface in APFP which can generate excess hydroxyl radicals. In the case of HPFP the rate of regeneration of Fe^{2+} ions is slow and hence its efficiency decreases at higher dye concentration. H_2O_2 serves as the better oxidizing agent in both the process over the entire range of dye concentration used. Hence APFP can be used as an efficient process for the degradation of the dye at higher concentrations compared to HPFP.

3.6 Role of scavenger on oxidizing agents

Hydroxyl radicals play a significant role in the oxidation of pollutants in the photo-Fenton process. To confirm the role of hydroxyl radicals in photo-Fenton's mechanism, the degradation process was carried out in the presence of hydroxyl radical scavengers like methyl alcohol. Methyl alcohol is able to deactivate hydroxyl radicals and their derivatives.²⁵ Methanol reacts with hydroxyl radicals with a second order rate constant of $9.7 \times 10^8 \text{ mol}^{-1} \text{ s}^{-1}$ as shown in eqn (20).

Though methanol is able to quench hydroxyl radicals, the extent of quenching differs for different oxidants and also for different processes. The decolorisation efficiency decreases from 100 to 32% for the process $\text{Fe}^0/\text{APS}/\text{UV}$ and to 15% for the process $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$. Similar drastic decreases in the rate constant is from 8.56×10^{-2} to $0.46 \times 10^{-2} \text{ min}^{-1}$ for H_2O_2 as an oxidant, while marginal decreases in the rate constant from 1.59×10^{-2} to $0.55 \times 10^{-2} \text{ min}^{-1}$ was observed for APS as oxidant for 15 min of UV illumination. This is because though hydroxyl radicals are quenched by the methanol, sulfate radicals still mediate the degradation process in the absence of hydroxyl radicals.²⁵ However for HPFP, decolorisation efficiency

decreased to 84% and 68% for the process $\text{Fe}^{2+}/\text{APS}/\text{UV}$ and $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$, respectively. The corresponding rate constants decreased from 0.87×10^{-2} to $0.56 \times 10^{-2} \text{ min}^{-1}$ for APS as oxidant and 1.21×10^{-2} to $0.51 \times 10^{-2} \text{ min}^{-1}$ for H_2O_2 as oxidant. By considering these changes in the magnitude of rate constant, it can be concluded that APFP generates excess hydroxyl radicals in the presence of H_2O_2 as an oxidant.

4.0 Kinetic studies

The calculation of apparent first order rate constant for APFP and HPFP at the time period of 16 and 90 min are shown in the Fig. 3 and 4. The dye solution was decolorised to an extent of 25%, 35% and 57% for the processes Fe^0/dark (A), $\text{Fe}^0/\text{APS}/\text{dark}$ (B) and $\text{Fe}^0/\text{H}_2\text{O}_2/\text{dark}$ (C) processes, respectively, in 16 min. The dye decolorisation in the presence of Fe^0/dark is due to the formation of iron complexes due to the reaction of iron ions with the dye molecules preferably with the chromophore of MR.²⁶ The lower efficiency of the dark process can be accounted due to the formation of stable iron complexes. The presence of oxidants like $\text{H}_2\text{O}_2/\text{APS}$, increases the decolorisation efficiency compared to the Fe^0/dark process. Iron powder under acidic conditions gets oxidized to Fe^{2+} ions, which further reacts with $\text{H}_2\text{O}_2/\text{APS}$ to generate hydroxyl radicals as shown in the eqn (8), (9) and (15). In the presence of oxidants alone (in the dark condition), dye decolorisation was almost negligible. For the process $\text{Fe}^{2+}/\text{dark}$ no significant dye decolorisation took place due to the formation of stable iron complexes in comparison with process (A). The azo chromophore can be reduced to amines on the iron surface even in dark conditions, which accounts for higher efficiency of Fe^0/dark compared to $\text{Fe}^{2+}/\text{dark}$ process.²⁷ Nearly 15–23% of the dye was decolorised for the process $\text{Fe}^{2+}/\text{APS}/\text{dark}$ and $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{dark}$ which suggests that both the oxidants showed comparable oxidation efficiency in the dark.

On UV irradiation for 15 min, the decolorisation efficiency for the processes A, B and C increased up to 32%, 48% and 100% respectively. The enhancement in the efficiency of Fe^0/UV process is due to the fact that iron complexes formed in the dark can be destroyed under UV irradiation and the iron ions can

Fig. 3 Plot of $-\log C/C_0$ versus time for various oxidation processes using Fe^0 as the catalyst. A. Fe^0/dark , B. Fe^0/APS , C. $\text{Fe}^0/\text{H}_2\text{O}_2$, D. Fe^0/UV , E. $\text{Fe}^0/\text{APS}/\text{UV}$, F. $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$.

Fig. 4 Plot of $-\log C/C_0$ versus time for various oxidation processes using Fe^{2+} as the catalyst. A. $\text{Fe}^{2+}/\text{APS}$, B. $\text{Fe}^{2+}/\text{H}_2\text{O}_2$, C. Fe^{2+}/UV , D. $\text{Fe}^{2+}/\text{APS}/\text{UV}$, E. $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$.

return to iron regeneration cycle for active participation in the cyclic photo-Fenton process.²⁸ The generation of hydroxyl radicals by the oxidants alone and the photolysis of dye molecules additionally contribute to the overall enhancement in the decolorisation process. In addition, photo reduction of Fe^{3+} to Fe^{2+} ions takes place at a faster rate on the iron surface in APFP²⁹ as shown in eqn (21).

The % decolorisation of the above mentioned dark homogeneous Fenton process increases up to 11, 68 and 100 % for 90 minutes of UV illumination respectively. The low decolorisation efficiency in the HPFP process can be accounted to the slow photo reduction process of ferric ions to ferrous ions.

The rate constant calculated for $\text{H}_2\text{O}_2/\text{UV}$ process is two times higher than APS/UV process. This difference can be attributed to the fact that the direct photolysis of H_2O_2 yields two hydroxyl radicals, while one hydroxyl radical is generated along with a proton by the APS as shown in eqn (15). Since the reaction was carried out in acidic conditions (pH 3), the additional protons generated lower the pH of the reaction medium and act as hydroxyl radical scavengers.

The rate constant calculated for the process $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$ is 3.75 times higher than that of $\text{Fe}^0/\text{H}_2\text{O}_2/\text{dark}$, while the rate constant for $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$ is ten times higher than that of $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{dark}$. The higher efficiency observed for both HPFP and APFP under illumination is mainly due to faster back photo reduction of Fe^{3+} ions to Fe^{2+} ions. The reduced Fe^{2+} ions can react with oxidants generating hydroxyl radicals thereby increasing the efficiency of the process. The rate constant calculated for the process $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV}$ is 7 times greater than $\text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV}$ due to the fact that Fe^{3+} ions so formed reduces to Fe^{2+} on the iron surface at a faster rate compared to homogeneous conditions.

The percentage of dye decolorisation can be compared for $\text{Fe}^0/\text{H}_2\text{O}_2/\text{dark}$ and $\text{Fe}^0/\text{APS}/\text{UV}$ process, both showing almost similar decolorisation efficiency. This is due to the fact that Fe^0 oxidizes to Fe^{2+} under acidic conditions (pH 3) and reacts with H_2O_2 generating hydroxyl radicals as shown in eqn (8). Further

reducing properties of the metallic iron surface enhances the process even in the dark.

The similarity in the rate of decolorisation for the process $\text{H}_2\text{O}_2/\text{UV}$ and Fe^0/dark can be accounted in the following way: the hydroxyl radicals generated from the direct photolysis of H_2O_2 attacks at the active site of azo chromophore yielding hydroxylated products resulting in decolorisation as shown in the Scheme 2. This kinetics of decolorisation are almost equal to the reducing nature of the iron surface which decolorises the dye yielding amine substituted products. These results suggest that iron powder is an efficient catalyst and can be used in heterogeneous photocatalysis. In all the above mentioned oxidation processes APFP process showed higher efficiency than their homogeneous counterparts.

The rate constant for the kinetics of decolorisation by various oxidation processes is of the order: $\text{Fe}^0/\text{H}_2\text{O}_2/\text{UV} > \text{Fe}^0/\text{H}_2\text{O}_2 > \text{Fe}^0/\text{APS}/\text{UV} > \text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{UV} > \text{Fe}^0/\text{UV} > \text{Fe}^0/\text{APS}/\text{dark} > \text{Fe}^0/\text{dark} \approx \text{H}_2\text{O}_2/\text{UV} > \text{Fe}^{2+}/\text{APS}/\text{UV} > \text{APS}/\text{UV} > \text{Fe}^{2+}/\text{H}_2\text{O}_2/\text{dark} > \text{Fe}^{2+}/\text{APS}/\text{dark} \approx \text{Fe}^{2+}/\text{UV}$. The process efficiency (Φ) implies the effectiveness of the present method. The process efficiency can be defined as the concentration of the pollutant degraded divided by the amount of energy in terms of intensity and exposure surface area time^{-1} .

Scheme 2 Probable degradation pathway for MR by HPFP.

Table 1 % decolorisation, rate constant and process efficiency for various oxidation processes^a

Oxidation processes	% decolorisation	Rate constant from $-\log C/C_0$ versus time plot $\times 10^{-2}/\text{min}^{-1}$	Process efficiency (Φ) $\times 10^{-12}/\text{ppm}$ Einstein ⁻¹
UV/APS	12	0.35	10.92
UV/H ₂ O ₂	23	0.70	20.93
Fe ⁰ /dark	25	0.74	22.75
Fe ²⁺ /dark	0	0.00	0
Fe ⁰ /APS/dark	35	0.96	31.85
Fe ⁰ /H ₂ O ₂ /dark	57	2.28	51.87
Fe ²⁺ /APS/dark	15	0.07	2.43
Fe ²⁺ /H ₂ O ₂ /dark	23	0.12	3.726
Fe ⁰ /UV	32	1.06	29.12
Fe ²⁺ /UV	11	0.05	1.782
Fe ⁰ /APS/UV	48	1.59	43.68
Fe ⁰ /H ₂ O ₂ /UV	100	8.56	91.00
Fe ²⁺ /APS/UV	68	0.56	11.01
Fe ²⁺ /H ₂ O ₂ /UV	100	1.21	16.20

^a For the dark process, unit of process efficiency is ppm.

$$\Phi = \frac{C_0 - C}{t \cdot I \cdot S} \quad (22)$$

C_0 is the initial concentration of the dye substrate and C is the concentration at time, t and $(C_0 - C)$ denotes the reduced concentration of the dye in ppm. I is the irradiation intensity [$I = \text{Einstein}/\text{m}^2 \text{ s} = 8.36 \times \lambda \text{ (nm)} \times \text{power (Watt)}$] (where λ is 370 nm and power is 125 W). 'S' denotes the solution irradiated plane surface area is 176 cm² and t represents the irradiation time in minutes. The performance and quantum efficiency of the process is characterized by the process efficiency. APFP shows higher process efficiency than their homogeneous counterparts (Table 1).

5.0 GC-MS analysis

5.1 Degradation by APFP

The mass spectrum recorded at the time of decolorisation of the dye for the process Fe⁰/H₂O₂/UV (16 min) shows m/z peaks at 136, 93 and 122 attributed to the formation of 4,4-*N,N*-dimethyl aniline (1), aniline (2) and benzoic acid (3). The intermediate (1) is formed due to the complete reduction of azo bonds to amines on the iron surface. The intermediate (2) is formed by the decarboxylation of adduct C₆H₄NH₂COOH, while the intermediate (3) is formed by the loss of -NH₂ as NH₂OH. The solution after 40 min of UV irradiation showed m/z peaks at 94, 110 corresponding to the formation of hydroxylated aromatic derivatives like phenol (4) and 1,4 dihydroxy benzene (5). The m/z peaks at 123 and 78 correspond to the formation of nitrobenzene (6) and benzene (7), respectively. The dealkylation of *N,N*-dimethyl group followed by the loss of amine group and subsequent hydroxylation leads to the formation of 1,4 dihydroxy benzene and phenol, respectively. The intermediate aniline (2) further undergoes oxidation to give nitrobenzene (6). The decarboxylation of intermediate (3) results in the formation of benzene (7). The solution after 45 min of irradiation shows only one m/z peak at 108 corresponding to the formation of hydroquinone (8). After one hour of irradiation mass spectrum shows m/z peaks at 46, 90 and 76 is due to the formation of simple

aliphatic carboxylic acids like formic acid (9), oxalic acid (10) and glyoxalic acid (11), respectively. Ring opening reaction of intermediate (8) results in the formation of the above intermediates. After 120 min of UV irradiation, the mass spectrum showed no characteristic peaks belonging to any functional group confirming the complete mineralization of the dye. Based on the intermediates, a probable degradation pathway has been proposed in Scheme 1.

5.2 Degradation by HPFP

The solution at the time of decolorisation shows two m/z peaks at 138 and 110 corresponding to the formation of 2,hydroxy benzoic acid (12) and 1,4 dihydroxy benzene (13). The attack of hydroxyl radicals at the site of the azo bond and terminal *N,N*-dimethyl chain results in the formation of above intermediates. The solution irradiated after 3.5 h shows two m/z peaks at 94 and 122 corresponding to the formation of phenol (14) and benzoic acid (15). Phenol might result from the decarboxylation of intermediate (13) and also due to the dehydroxylation of intermediate (12). The mass spectrum recorded at 4 h of irradiation shows an intense m/z peak at 78 corresponding to the formation of benzene (16). The dehydroxylation and decarboxylation of the intermediates (14) and (15) might lead to the formation of benzene. After 4.5 h of irradiation no characteristic peaks were found in mass spectrum. Based on the intermediates, a probable degradation pathway is proposed in Scheme 2.

Conclusion

The degradation of MR was investigated by APFP and HPFP using H₂O₂ and APS as oxidants. APFP showed higher process efficiency and rate constant than their homogeneous counterparts. H₂O₂ served as a better oxidant for both the oxidation processes. The reactivity of the APS is lower probably due to the excess generation of protons which showed negative effect on the degradation rate by scavenging hydroxyl radicals. However, APS effectively inhibited the precipitation of iron by providing excess acidity to the reaction medium which is an essential condition for Fenton's process. The higher efficiency of APFP is attributed to the faster reduction of Fe³⁺ to Fe²⁺ ions on the iron surface. The

degradation pathway concluded that the initial mechanism in the APFP is the reduction of azo bonds to amines while in the case of HPFP it is the oxidation of azo bonds to hydroxylated intermediates. With optimization, the degradation using APFP can be a novel technology for the degradation of azo dye.

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